

NEWS AND VIEWS

COMMENT

Species–genetic diversity correlations in habitat fragmentation can be biased by small sample sizes

ALISON G. NAZARENO* and ALISTAIR S. JUMP†

*Federal University of Santa Catarina, CP 476, 88040-900 Florianópolis, Santa Catarina, Brazil, †Biological and Environmental Sciences, School of Natural Sciences, University of Stirling, Stirling FK9 4LA, UK

Abstract

Predicted parallel impacts of habitat fragmentation on genes and species lie at the core of conservation biology, yet tests of this rule are rare. In a recent article in *Ecology Letters*, Struebig *et al.* (2011) report that declining genetic diversity accompanies declining species diversity in tropical forest fragments. However, this study estimates diversity in many populations through extrapolation from very small sample sizes. Using the data of this recent work, we show that results estimated from the smallest sample sizes drive the species–genetic diversity correlation (SGDC), owing to a false-positive association between habitat fragmentation and loss of genetic diversity. Small sample sizes are a persistent problem in habitat fragmentation studies, the results of which often do not fit simple theoretical models. It is essential, therefore, that data assessing the proposed SGDC are sufficient in order that conclusions be robust.

Keywords: genetic diversity, population genetics, rarefaction method, sampling strategy, small sample size

Received 1 December 2011; revision received 3 February 2012; accepted 20 February 2012

Understanding the impacts of habitat fragmentation on genes and species is fundamentally important to conservation biology and landscape ecology. However, there is a broad consensus that the genetic consequences of habitat fragmentation are more complex than those predicted by simple theoretical models (Ewers & Didham 2006; Feeley & Terborgh 2008; Bacles & Jump 2011). Because of the dependence of responses on a species' ecology and evolutionary history, it has been difficult to quantitatively summarize

results reported to date in order to reach sound conclusions about the effects of habitat fragmentation on population genetics.

Methodological limitations of habitat fragmentation surveys can also weaken our ability to understand ecological mechanisms underlying forest fragmentation effects. For example, sample size, replicate independence and uncontrolled variation in confounding variables remain critical problems in fragmentation studies (McGarigal & Cushman 2002). Consequently, understanding the impacts of habitat fragmentation on genes and species and the relationship between the two—the species–genetic diversity correlation (SGDC)—poses a major challenge in data collection if reliable conclusions are to be reached. To illustrate this point, we explore the impact of small sample sizes in a recent test of the SGDC in forest fragments by Struebig *et al.* (2011).

In an overall excellent study, Struebig *et al.* (2011) compare parallel impacts of habitat fragmentation on species and genetic diversity of insectivorous bats. The authors attempted to synthesize the SGDC across species with contrasting ecological traits using richness-based measures. Taken at face value, their results indicate that an area-dependent decrease in species richness is mirrored by a parallel decrease in genetic diversity.

However, Struebig *et al.* (2011) report a negative impact of fragmentation on the genetic diversity of populations of just one (*Kerivoula papillosa*) of three bat species studied. The authors suggest that the inability to detect a correlation between fragment size and the genetic diversity of the other two species studied is heavily dependent on their ecological traits. Although this explanation may well be true, it is equally important to point out that results for this species have been estimated via rarefaction methods based on very small sample sizes and extrapolated to a common sample size of 15 (whilst 63% of the population samples include <15 individuals, 19% of samples include five individuals or less). Results estimated from the smallest sample sizes are likely to be highly error prone and may, therefore, produce false positives in the final analysis. Indeed, small sample sizes are *per se* the greatest problem in studies that aim to detect genetic consequences of habitat fragmentation, and there is no adequate current method to correct for low sampling intensity (Belkhir *et al.* 2006).

Whilst Struebig *et al.* (2011) deserve credit for an elegant study, their methods highlight intrinsic problems in analyses that are substantially based on small sample sizes. The problem arises from the fact that the authors used hyper-variable loci (e.g. microsatellite markers) from very small samples. Such loci are acknowledged as ineffective in the analysis of small sample sizes, when they provide very little information about the presence of rare alleles (Leberg 2002; Belkhir *et al.* 2006).

Correspondence: Alison G. Nazareno, Fax: +55 48 37215322; E-mail: alison_nazareno@yahoo.com.br

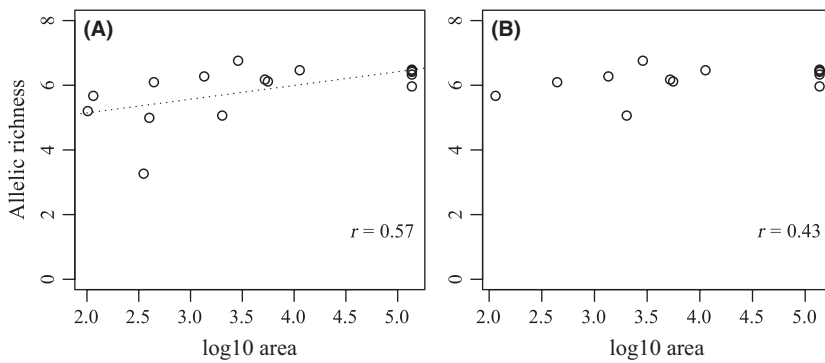


Fig. 1 Effects of small sample size on relationship between allelic richness and log₁₀ area of *Kerivoula papillosa* populations: (A) negative impact of forest fragmentation on the genetic diversity with sample sizes larger than three individuals ($P = 0.023$); (B) no impact of forest fragmentation with sample sizes larger than five individuals ($P = 0.146$).

Allelic richness is highly impacted by population reduction because of the rapid elimination of rare alleles (Cornuet & Luikart 1996). Consequently, it is unclear how Struebig *et al.* (2011) can detect a loss of rare alleles from populations based on data estimated from only a few individuals. For instance, the *Kp30* locus used by the authors in the genetic analysis of *K. papillosa* shows 19 alleles across a sample of 22 individuals (Struebig *et al.* 2008). In this case, the apparent reduction in allelic richness observed for *K. papillosa* by Struebig *et al.* (2011) cannot reliably be attributed to forest fragmentation because the sample size analysed was not adequate to detect rare alleles and was, therefore, insufficient to assess their loss. Critically, although the rarefaction method they employ returns a reassuringly narrow confidence interval on allelic richness estimated from very small population samples, those samples are likely to be a very poor representation of the population from which they are derived (Leberg 2002). With issues as complex as habitat fragmentation and the SGDC, no comparable study to Struebig *et al.*'s (2011) study has used such limited samples (Vellend 2004; Cleary *et al.* 2006; He *et al.* 2008; Evanno *et al.* 2009).

To explore the effect of small sample size on the results reported for *K. papillosa* by Struebig *et al.* (2011), we used the Pearson correlation coefficient between allelic richness and log₁₀ area from data presented by Struebig *et al.* (2011), log₁₀ transforming area due to reduced heteroscedasticity. Like Struebig *et al.* (2011), we find a significant correlation between variables ($N = 16$, Pearson $r = 0.57$, $P = 0.023$, see Fig. 1A). However, when samples with <6 individuals are excluded from the analysis, there is no significant correlation between habitat area and the allelic richness (fragments and continuous forest samples: $N = 13$, Pearson $r = 0.43$, $P = 0.146$, see Fig. 1B; fragments only: $N = 8$, Pearson $r = 0.39$, $P = 0.341$). Furthermore, the relationship between mean number of alleles per locus and sample size becomes stronger when small samples are excluded (rising from $r = 0.64$ with all samples to $r = 0.84$ with sample sizes ≥ 10 ; both significant at $P = 0.05$), suggesting overall that the results observed by Struebig *et al.* (2011) for *K. papillosa* are significantly biased by the very small sample sizes included in this study. In effect, the positive SGDC reported appears to be driven by only three small population samples in a single species.

As reduction in population sizes will become more common because of habitat fragmentation and degradation, it is paramount that we understand the impact of habitat reduction on diversity both within and between species. However, it is equally important that data assessing the relationship between these measures are sufficient in order that conclusions be robust. Whilst the work by Struebig *et al.* (2011) remains highly valuable overall in its contribution to conservation biology, it does not lend strong support to the proposed SGDC. Further work based on greater sample sizes and additional measures, such as the reduction in genetically effective population size (N_e), is required to adequately assess this rule.

Acknowledgements

AGN was supported by a doctoral fellowship of CNPq (National Council of Technological and Scientific Development, Brazil).

References

- Bacles CFE, Jump AS (2011) Taking a tree's perspective on forest fragmentation genetics. *Trends in Plant Science*, **16**, 13–18.
- Belkhir K, Dawson KJ, Bonhomme F (2006) A comparison of rarefaction and bayesian methods for predicting the allelic richness of future samples on the basis of currently available samples. *Journal of Heredity*, **97**, 483–492.
- Cleary DFR, Fauvelot C, Genner MJ, Menken SBJ, Mooers A (2006) Parallel responses of species and genetic diversity to El Niño Southern Oscillation-induced environmental destruction. *Ecology Letters*, **9**, 304–310.
- Cornuet JM, Luikart G (1996) Description and power analysis of two tests for inferring recent population bottlenecks from allele frequency data. *Genetics*, **144**, 2001–2014.
- Evanno G, Castella E, Antoine C, Paillat G, Goudet J (2009) Parallel changes in genetic diversity and species diversity following a natural disturbance. *Molecular Ecology*, **18**, 1137–1144.
- Ewers RM, Didham RK (2006) Confounding factors in the detection of species responses to habitat fragmentation. *Biological Reviews*, **81**, 117–142.
- Feeley KJ, Terborgh JW (2008) Direct vs. indirect effects of habitat reduction on the loss of avian species from tropical forest fragments. *Animal Conservation*, **11**, 353–360.
- He T, Lamont BB, Krauss SL, Enright NJ, Miller BP (2008) Covariation between intraspecific genetic diversity and species diversity within a plant functional group. *Journal of Ecology*, **96**, 956–961.

- Leberg PL (2002) Estimating allelic richness: effects of sample size and bottlenecks. *Molecular Ecology*, **11**, 2445–2449.
- McGarigal K, Cushman SA (2002) Comparative evaluation of experimental approaches to the study of habitat fragmentation effects. *Ecological Applications*, **12**, 335–345.
- Struebig MJ, Horsburgh G, Pandhal J *et al.* (2008) Isolation and characterisation of microsatellite loci in the papillose woolly bat, *Kerivoula papillosa* (Chiroptera: Vespertilionidae). *Conservation Genetics*, **9**, 751–756.
- Struebig M, Kingston T, Petit EJ, Zubaid A, Mohd-Adnan A, Rossiter SJ (2011) Parallel declines in species and genetic diversity in tropical forest fragments. *Ecology Letters*, **14**, 582–590.
- Vellend M (2004) Parallel effects of land-use history on species diversity and genetic diversity of forest herbs. *Ecology*, **85**, 3043–3055.

A.G.N. has interests in evolutionary ecology, population genetics and conservation of threatened tree species. A.S.J. is interested in impacts of environmental change on plant diversity, structure, and function, from populations to species distributions.

doi: 10.1111/j.1365-294X.2012.05611.x